Theory-Driven Analysis of Ecological Data



Day 3: Spatial data

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Outline of the day

Morning:

- Introduction : typical data? typical questions?
- Part 1: Descriptive approaches (aka statistical models)
- Part 2: Process-based approaches (aka mechanistic models)
- **Conclusion** : what's best, if anything?

Afternoon :

- **Practical:** simulating metacommunity data, fitting a patch occupancy model to infer competitive interactions

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Evening speaker : Davide Martinetti, INRAE, Avignon

"Predicting the risk of invasion of the Japanese beetle *Popillia japonica* in Europe".

A set of species occurring at the same place, at the same time

A set of species occurring at the same place, at the same time

Spatial boundaries can be arbitrary (study plots)

Plants on forest

Invertebrates in grassland

Corals on reefs

A set of species occurring at the same place, at the same time

Snails in ponds

Ants on islands

Flies on fruits

A set of species occurring at the same place, at the same time

Not restricted to one trophic level or resource use

Plants & herbivores on forest

Invertebrates & plants in grassland

Corals & fishes on reefs

The typical spatial dataset

The core thing

Species (S)

Community matrix

Sampling units

Presence/Absence (Detection/Non Detection)

Abundance

Detection only

The typical spatial dataset Env. var. (P) The core thing Sampling units (N) Species (S) Environmental variables Community matrix Presence/Absence Sampling units (Detection/Non Detection) Coordinates Abundance Sampling units (N) Detection only Spatial positions

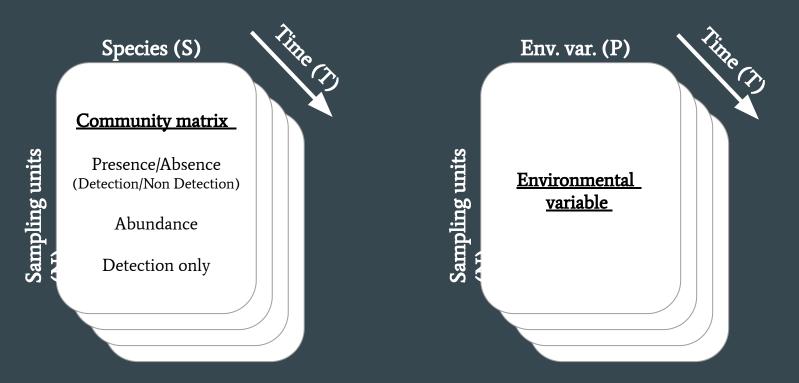
Additional data

The typical spatial dataset Additional data Env. var. (P) The core thing Species (S) Sampling units (N) Species (S) Species (S) Environmental Phylogeny variables Community matrix Presence/Absence (Detection/Non Detection) Coordinates Abundance Traits (T) Sampling units (N) Detection only Species (S) Spatial <u>Sp. traits</u> positions

Sampling units

The typical spatial dataset

And observations can be repeated in time



The typical spatial dataset

And observations can be repeated in time

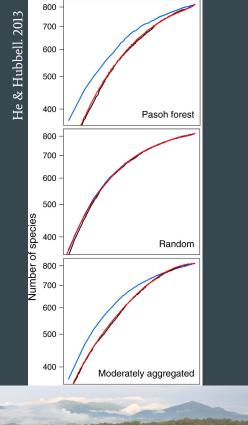
Note: having temporally-repeated can greatly improve things

- we can avoid assumptions about system equilibrium (species dyn. and/or envt/species relationships)
- it can help identify interactions among species
- we can more properly account for species detection/non detection (short-term repeats)

The typical questions addressed

- Patterns of ecological diversity
 - Species-area relationships (SAR curves)
 - Distribution of species richnesses (α , γ)
 - How similar / dissimilar communities are? (β)

Pasoh Nature Reserve of Malaysia



0.2 0.5 2.0 5.0 20.0 50.0 Area (ha)

The typical questions addressed

• Patterns of ecological diversity

- Species-area relationships (SAR curves)
- Distribution of species richnesses (α , γ)
- How similar / dissimilar communities are? (β)

• Species distributions

- Habitat preferences / specialization / generalism?
- Are species interacting, and if so, how?
- How important is dispersal versus habitat or competition?
- How contingent is community assembly?
- ... many more

The typical questions addressed

• Patterns of ecological diversity

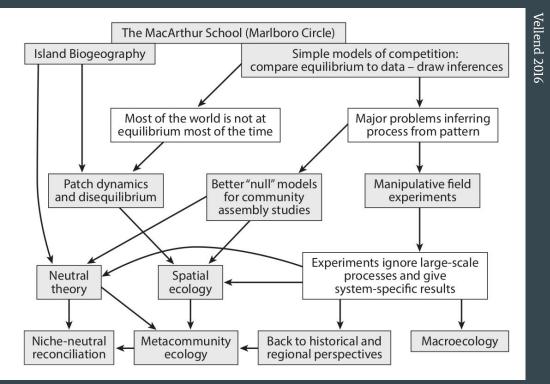
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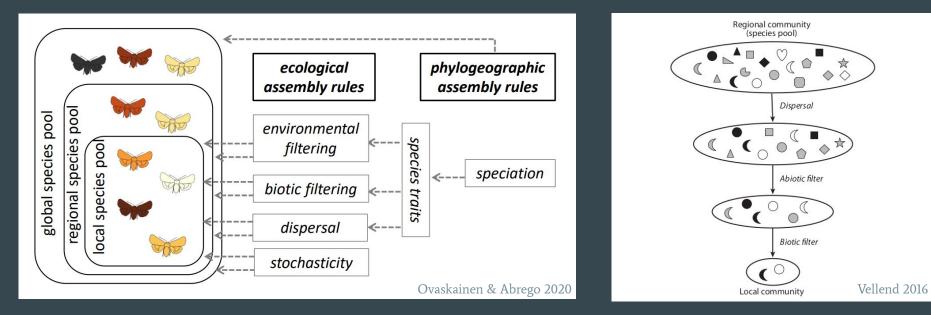
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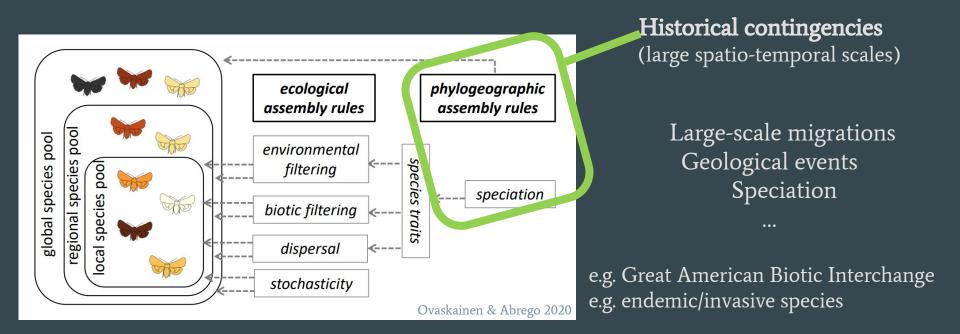
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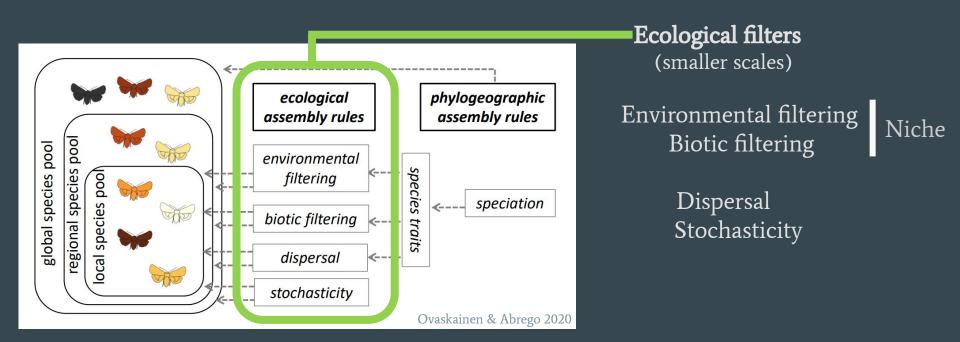
• Theoretical frameworks to guide analyses?

• We'll make a long story short...



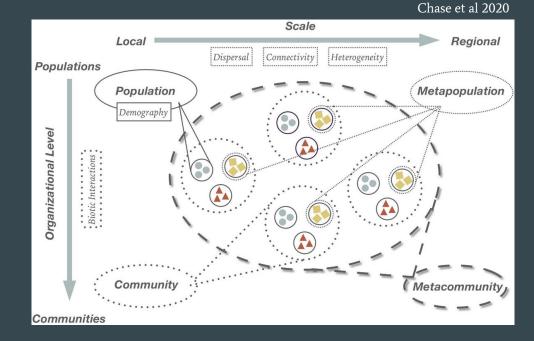




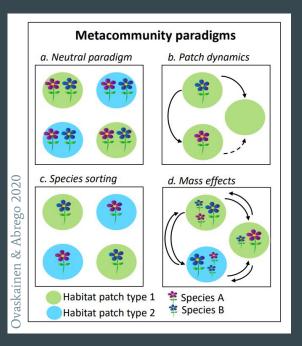


- A. The 'assembly rules' framework
- Strengths
 - Very intuitive and broadly applicable
- Limitations
 - Mostly a static description, with a hierarchical set of filters
 - Focus on each individual site, not on the coupling between them

- B. The metacommunity framework
- A metacommunity is set of interconnected local communities
- Directly follows from the metapopulation paradigm



B. The metacommunity framework



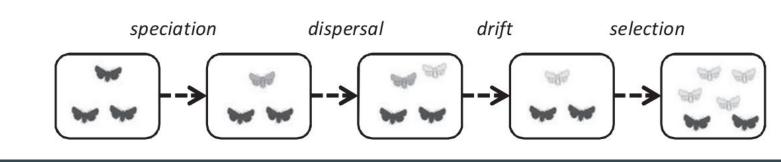
	(A)	Spatial environment			
	Species attributes	Homogenous	Heterogeneous		
	Different niches	1. Patch dynamics (see Chapter 13)	 2. Mass effects/source–sink dynamics 3. Species sorting 		
	Demographically equivalent	4. Neutral theory			
Mittelbach & McGill 2019	Consequences of (B) patch heterogeneity	PD SS PD SS With Comp- Col NM Connectivity/dispersal	ME		

- B. The metacommunity framework
- Strengths
 - Emphasizes dynamical processes and the coupling of communities at different scales
 - Reunified niche, spatial and neutral perspectives

- Limitations
 - A collection of models that differ along many axes
 - The four paradigms are not mutually exclusive, and do not map to different processes.
 They are hard to disentangle

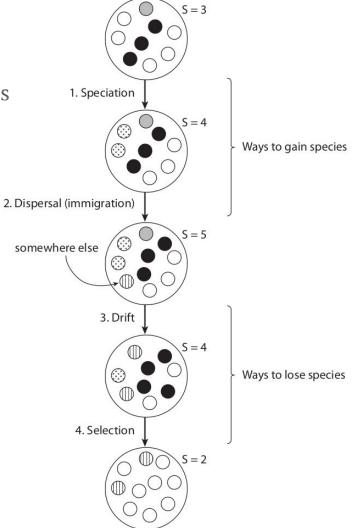
C. Vellends' theory of ecological communities

• Vellend's theory recycles the framework of population genetics (and its four evolutionary 'forces'):



Ovaskainen & Abrego 2020

C. Vellends' theory of ecological communities



Vellend 2016

- Vellends' theory of ecological communities C.
- Strengths
 - Emphasizes true and distinct dynamical processes Ο
 - Proximity with population genetics and evolutionary theory Ο

- Limitations
 - Order of the processes is arbitrary. 0
 - 15/revisiting-vellend-2010/ Speciation probably acts on a different timescale (same issue with Neuron 0

https://reflectior

The theory is probably too abstract/generic to be very operational. 0

Theory-driven data analysis

So, in practice: what analyses?

• Part 1: descriptive approaches (more related to assembly rules and metacommunity frameworks)

• Part 2: process-based approaches (more related to metacommunity and Vellend's frameworks)

Bridge intro / part 1

• questions on Intro



• Statistical approaches (community level) - multivariate analyses (think PCA)

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 - (Unconstrained) ordination methods (CA, PCoA, MDMS)
 - multivariate similarity among species / patches,
 - main axis can be (linearly or not) linked to environmental gradients or traits a posteriori
 - (Constrained) ordination methods (CCA, RDA, dbRDA, ...)
 - environmental data / spatial positions are incorporated to constrained the ordination
 - variance can be partitioned among multiple explicative tables (e.g. envt vs. space)

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Based on the notion of a distances among sampling units / species.

Numerous (dis)similarity measures: Euclidean, Bray-Curtis, Chi Square or Euclidean dist. on transformed data (Profiles, Hellinger, ...)

Numerous ways of representing space: spatial neighborhood, spat. weighting matrices (SWM), spatial predictors (orth. polynomials of coordinates, PCNMs, MEMs, etc.)

e.g. Guadeloupe Mollusks

D. surinamense









P. glauca

P. acuta (1972)



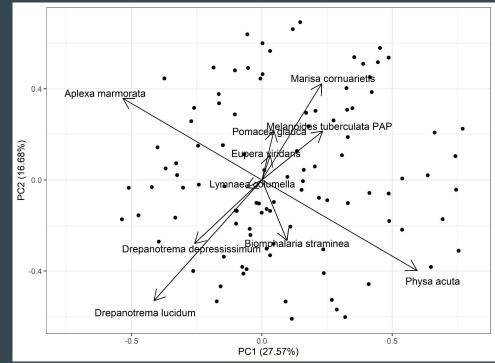


P. columella (1997)

M. tuberculata (1979)

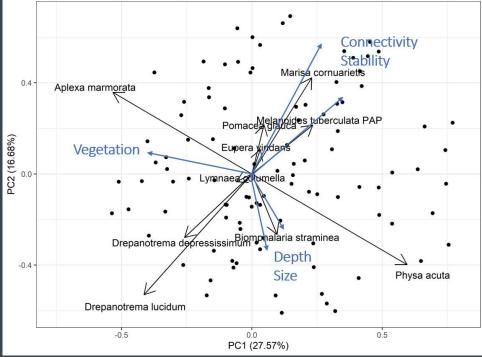


e.g. Guadeloupe Mollusks: unconstrained (PCoA)



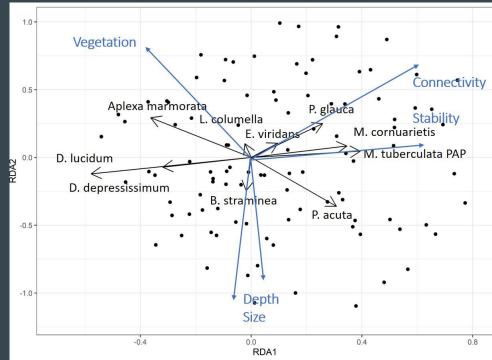
e.g. Guadeloupe Mollusks: unconstrained (PCoA) *A posteriori* environmental fit

	Dim1	Dim2	r2	Pr(>r)	
Vegetation	-0.97466	0.22371	0.1167	0.001	***
Size	0.15422	-0.98804	0.0698	0.001	કેર કેર કેર
Depth	0.39564	-0.91841	0.0521	0.009	\$t \$t
Stability	0.71490	0.69923	0.1207	0.001	* * *
Connectivity	0.44110	0.89746	0.2272	0.001	sie sie sie



e.g. Guadeloupe Mollusks: RDA (constrained)

	Inertia	Proportion	Rank
Total	0.5950		
Constrained	0.0868	0.1459	5
Unconstrained	0.5082	0.8541	10
Inertia is van	riance		



e.g. Guadeloupe Mollusks: partial-RDA & variance partitioning with PCNMs Widely used after Cottenie (2005) publication

the idea, partitioning variance between environmental and spatial predictors

as a way to distinguish whether communities result from niche-based or neutral-based processes...

terms of the validity of neutral model. I collected 158 published data sets with

	Ecology Letters, (2005) 8: 1175-1182	doi: 10.1111/j.1461-0248.2005.00820.x				
LETTER	Integrating environmental and spatial processes in ecological community dynamics					
Ĵ.	Abstract					
Karl Cottenie National Center for Ecological Analysis and Synthesis,	cornerstone of modern ecology. In these	s of species across multiple sites form the metacommunities, the relative importance of cesses is currently hotly debated, especially in				

University of California, Santa

e.g. Guadeloupe Mollusks, partial-RDA & **variance partitioning** with PCNMs Widely used after Cottenie (2005) publication

the idea, partitioning variance between environmental and spatial predictors

[E]	[S]	[E S]	[S E]	Metacommunity type
(not) sig.	(not) sig.	sig.	not sig.	SS
(not) sig.	(not) sig.	sig.	sig.	SS + ME
(not) sig.	(not) sig.	not sig.	sig.	NM/PD
(not) sig.	(not) sig.	not sig.	not sig.	Undetermined
not sig.	not sig.	not sig.	not sig.	No found

Table 1 Decision tree for relationship between significance struc-

ture and metacommunity types

Relationship between significance structure of the four important variation components and associated metacommunity types. The components are environment [E], space [S], environment independent of space $[E \mid S]$, and space independent of environment $[S \mid E]$. sig., variation component explains a significant part of the variation in community structure; not sig., no significant part; (not) sig., either.

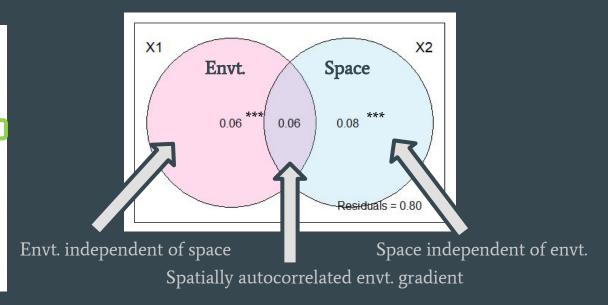
Deeply rooted in the metacommunity framework

e.g. Guadeloupe Mollusks, partial-RDA & variance partitioning with PCNMs

Using Principal Coordinates of Neighbourhood Matrix (PCNM, Borcard & Legendre, 2002) as spatial descriptors

[E]	[S]	[E S]	[S E]	Metacommunity type
(not) sig.	(not) sig.	sig.	not sig.	SS
(not) sig.	(not) sig.	sig.	sig.	SS + ME
(not) sig. (not) sig.	(not) sig. (not) sig.	not sig. not sig.	sig. not sig.	Undetermined
not sig.	not sig.	not sig.	not sig.	No found

Relationship between significance structure of the four important variation components and associated metacommunity types. The components are environment [E], space [S], environment independent of space [E | S], and space independent of environment [S | E]. sig., variation component explains a significant part of the variation in community structure; not sig., no significant part; (not) sig., either.



- Statistical approaches (community level) multivariate analyses
 - Among distance matrices regressions
 - Mantel and partial Mantel tests
 - MRM (Lichstein, 2007)

Might be better suited for testing some theories (e.g. Neutral Theory), See Tuomisto & Ruokolainen, 2006.

But ... see also Tuomisto, Ruokolainen & Ruokolainen, 2012.

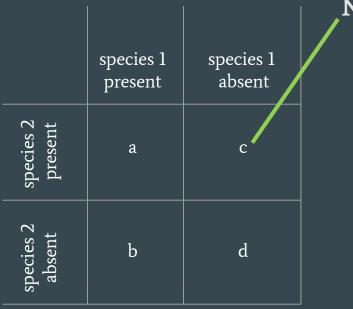


• Statistical approaches (species level)

Rationale : species that do not interact and have the same habitat affiliations should be distributed **independently** over sites*.

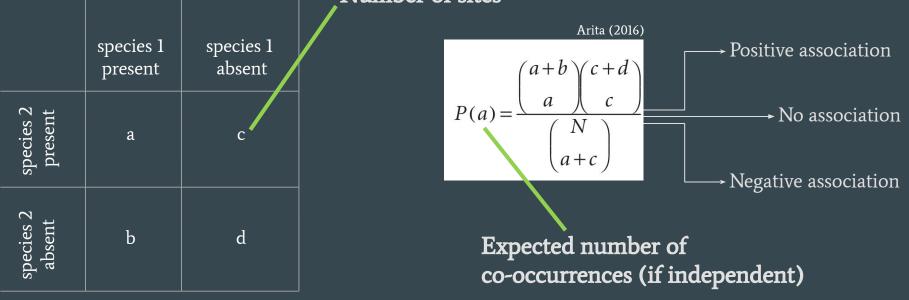
*Think it through

- Statistical approaches (species level)
 - \sim Fitting null distributions, pairwise tests of species independence (Forbes 1907, Veech et al . 2013)



Number of sites

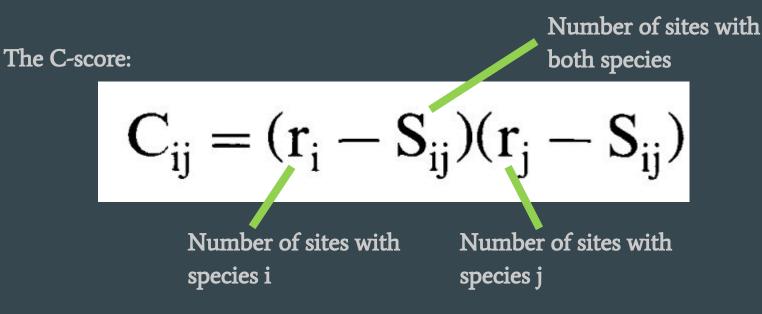
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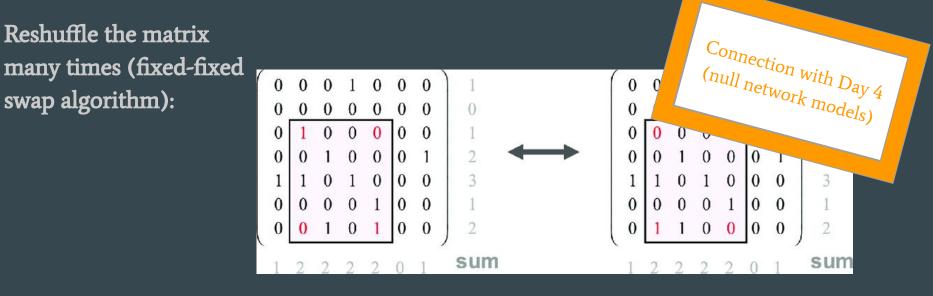
Number of sites

- Statistical approaches (species level)
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 - Permutation based (null model) approaches (Diamond, Simberloff, Gotelli, Ulrich)

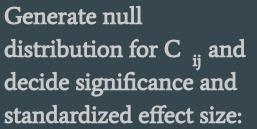
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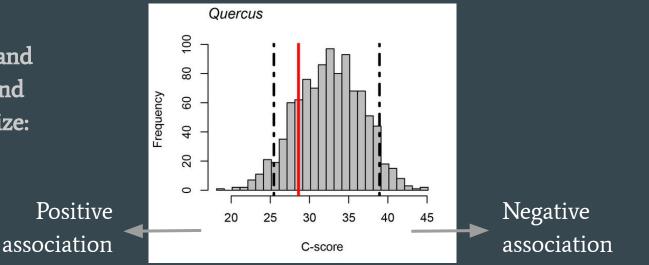


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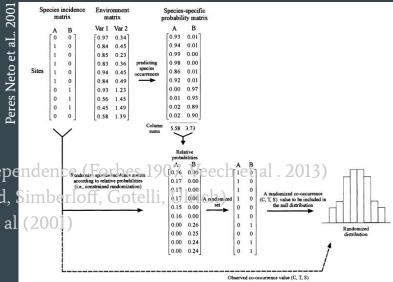
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 - Extension to constrained null models by Peres Neto et al (2001)

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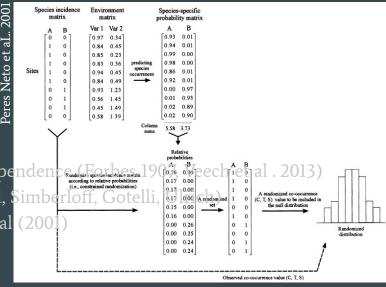


Table 5. Pair-wise associations between hunting spiders. Positive associations (+) were assessed by the significance of the T-score (note: S-score provided similar results). Negative associations (-) were judged by the significance of the C-score. All results based on alpha = 0.05. The upper diagonal contains the results based on the unconstrained null model, whereas the lower diagonal has the results for Cr-RA1 (note: Cr-RA2 provided similar results). Species codes follow Table 2.

	AL	PL	ZS	PN	PP	AA	TT	AC	PM	AAc	AF	AP
AL			+	+	+	+		+	8 CH - 18 H			
PL								+		_		-
ZS				+	+	+		+		_		_
PN					+	+		+			-	
PP						+		+	+			_
AA					+			+	+			-
TT AC												_
AC				+	+	+						1
PM												
AAc												
AF												
AP												

- Statistical approaches (species level)
 - (Joint) Species Distribution Modelling
 - Occupancy models (emphasis on detection process)

- Statistical approaches (species level): Species Distribution Models (SDM)
 - From a theoretical viewpoint : they catch the 'realized niche' of one species through regression or a broad range of statistical/classification methods
 - Main aim : making predictions on past, contemporary, and future sp. distributions
 - \circ Have been criticized regarding the lack of consideration for :
 - equilibrium assumption (species are wherever they can) (Araújo & Pearson, 2005)
 - dispersal, dispersal limitation in particular
 - species interactions (but see Anderson, 2017, JoB)
 - but also sampling, methods, model transferability...

• Statistical approaches (species level): Species Distribution Models (SDM)

A **lot** of methods...

Table 3.1 Summary of some popular and recently emerged SDM frameworks used to model community data. The SDM frameworks are classified either as single-species distribution models or joint species distribution models.

Single-species distribution models	Reference				
Boosted regression trees (BRT)	Hijmans et al. (2017); Ridgeway (2017)				
Generalised additive model (GAM)	Wood (2011)				
Generalised linear model (GLM)	R Development Core Team (2019)				
Gradient nearest neighbour (GNN)	Crookston & Finley (2008)				
Maximum-entropy approach (MaxEnt)	Phillips et al. (2006)				
Multivariate adaptive regression spline (MAR.S-COMM)	Milborrow (2017)				
Multivariate regression tree (MR TS)	De'ath et al. (2014)				
Random forest (RF)	Liaw & Wiener (2002)				
Support vector machine (SVM)	Meyer et al. (2017)				
Gradient extreme boosting (XGB)	Chen et al. (2018)				

From Ovaskainen & Abrego, 2020

• Statistical approaches (species level): Species Distribution Models (SDM)

With several species:

- One SDM per species, then combine —> **Stacked** SDMs
- SDM for all species at the same time —> **Joint** SDMs (JSDMs)

Think: several point estimates of means versus one ANOVA

• Statistical approaches (species level): Species Distribution Models (SDM)

Even more methods...

Joint species distribution models					
Bayesian community ecology analysis (BC)	Golding & Harris (2015)				
Bayesian ordination and regression analysis (BORAL)	Hui (2017)				
Generalised joint attribute modelling (GJAM)	Clark et al. (2017)				
Hierarchical modelling of species communities (HMSC)	Ovaskainen et al. (2017b)				
Multivariate stochastic neural network (MISTN)	Harris (2015)				
Species archetype model (SAM)	Hui et al. (2013)				

From Ovaskainen & Abrego, 2020

- JSDMs: one relatively simple and flexible approach
 - Hierarchical Models of Species Composition (HMSC) Ovaskainen et al. (2017)

- Bayesian framework
- Several latent factors (random effects)
- Species occupancies + habitat variables + species traits + Phylogeny +

- JSDMs: one relatively simple and flexible approach
 - Hierarchical Models of Species Composition (HMSC) Ovaskainen et al. (2017)

Nice R package, nice book,

Methods in Ecology and E

APPLICATION 🖻 Open Access 💿 🛈 😒

New Results

A Follow this preprint

Accelerating joint species distribution modeling with Hmsc-HPC: A 1000x faster GPU deployment

Anis Ur Rahman, Gleb Tikhonov, Jari Oksanen, Tuomas Rossi, Otso Ovaskainen doi: https://doi.org/10.1101/2024.02.13.580046

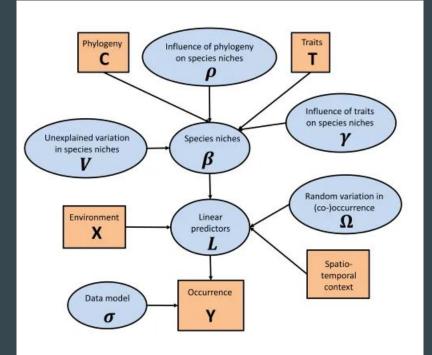
This article is a preprint and has not been certified by peer review [what does this mean?].

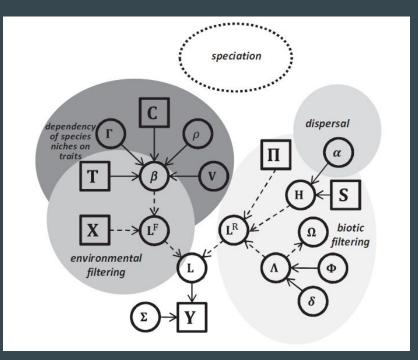
Joint species distribution modelling with the R-package HMSC

Gleb Tikhonov, Øystein H. Opedal, Nerea Abrego, Aleksi Lehikoinen, Melinda M. J. de Jonge, Jari Oksanen, Otso Ovaskainen 🗙

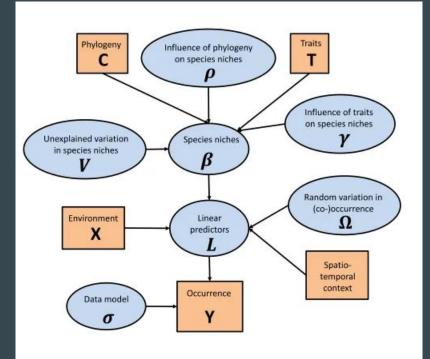
First published: 25 December 2019 | https://doi.org/10.1111/2041-210X.13345 | Citations: 5

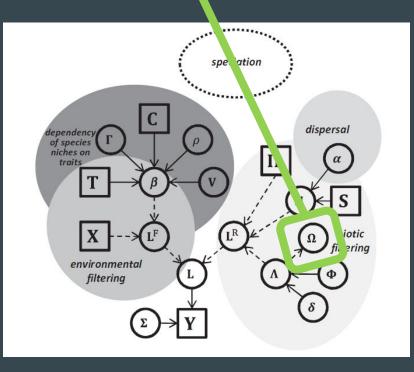






Species interactions must be deduced from **residual covariances***





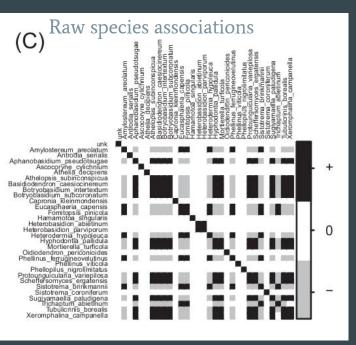
*Same rationale as for null-model approach earlier

• An example of HMSC application: fungal communities of decaying trees



Botryobasidium subcoronatum (@Wikipedia)

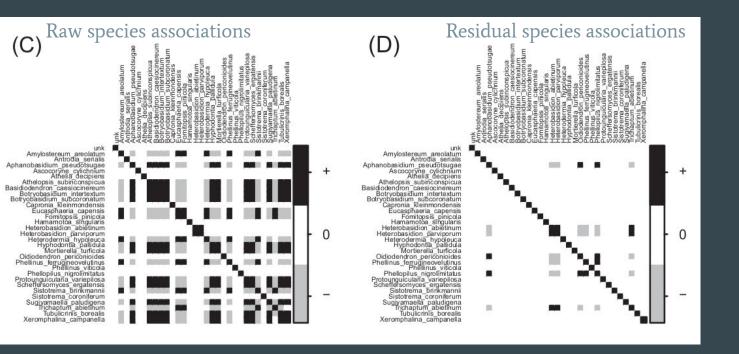
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Botryobasidium subcoronatum (@Wikipedia)

• An example of HMSC application: fungal communities of decaying trees



From Ovaskainen & Abrego, 2020

• Statistical approaches (species level): Occupancy models



Ecography 40: 281–295, 2017 doi: 10.1111/ecog.02445 © 2016 The Author. Ecography © 2016 Nordic Society Oikos Subject Editor: Miguel Araújo. Accepted 15 June 2016

Modelling of species distributions, range dynamics and communities under imperfect detection: advances, challenges and opportunities

Gurutzeta Guillera-Arroita

G. Guillera-Arroita (gurutzeta.guillera@unimelb.edu.au), School of Biosciences, Univ. of Melbourne, Australia.

Building useful models of species distributions requires attention to several important issues, one being imperfect detection of species. Data sets of species detections are likely to suffer from false absence records. Depending on the type of survey, false positive records can also be a problem. Disregarding these observation errors may lead to important biases in model estimation as well as overconfidence about precision. The severity of the problem depends on the intensity of these errors and how they correlate with environmental characteristics (e.g. where species detectability strongly depends on habitat

• Statistical approaches (species level): Occupancy models

Close to SDMs, but account for imperfect species detection (non-detection ≠ absence) The idea: distinguish *true* occupancy states from *observed* occupancy states, and estimate a **detection probability** which allows to account for false negatives

Use short-term survey repeats (short enough to assume no change in occupancy state)

• Statistical approaches (species level): Occupancy models

Close to SDMs, but account for imperfect species detection (non-detection ≠ absence)
 The idea: distinguish *true* occupancy states from *observed* occupancy states,
 and estimate a **detection probability** which allows to account for false negatives

Bufo americanus



Initial survey: detected in *ca.* 30% of sites

Considering detection probability: probable presence in *ca.* 50% (+44%) of sites

MacKenzie et al. 2002

• Statistical approaches (species level): Occupancy models

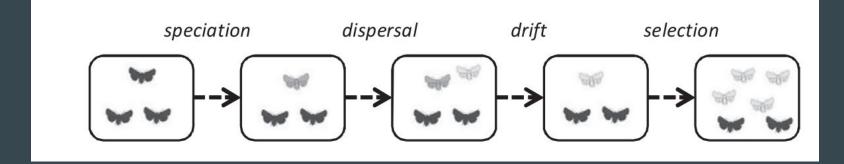


Bridge part 1 / part 2

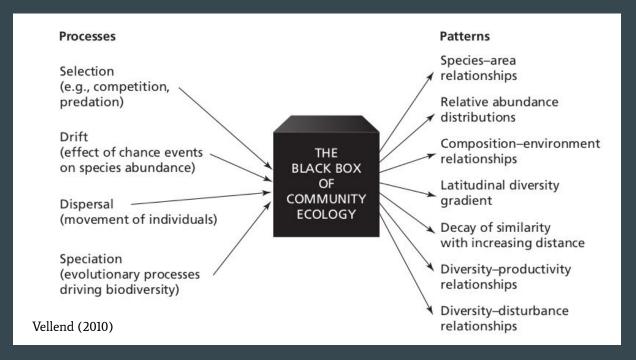
• Coffee break / questions on Part 1



• Describing the processes and how they operate in space and time

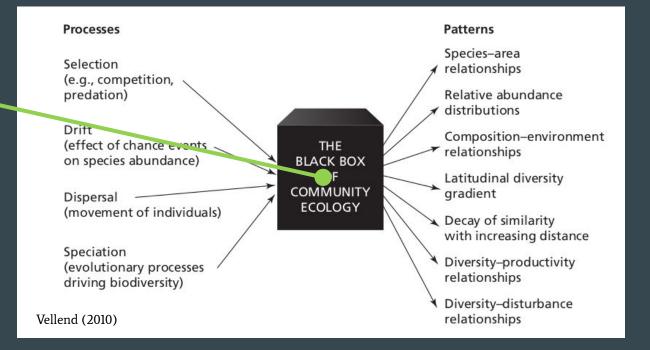


• Describing the processes and how they operate in space and time



• Describing the processes and how they operate in space and time

Some mathematical (dynamical) model —



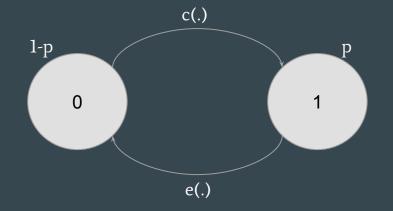
• A. Patch occupancy models

In one site (patch), a species is either present (p) or absent (1-p)

Local dynamics can be neglected.

The probability of presence is a dynamic equilibrium between:

- The rate of colonization (c)
- The rate of extinction (e)



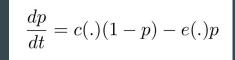
• A. Patch occupancy models

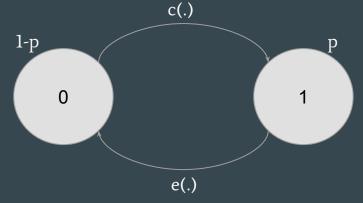
In one site (patch), a species is either present (p) or absent (1-p)

Local dynamics can be neglected.

The probability of presence is a dynamic equilibrium between:

- The rate of colonization (c)
- The rate of extinction (e)





• A. Patch occupancy models

$$\frac{dp}{dt} = c(.)(1-p) - e(.)p$$

Now if you assume that:

- The extinction rate e(.) is just a constant: e(.) = e
- The colonization rate is proportional to the overall occupancy, which, if all patches are homogeneous, is just p: c(.) = c p

you get

$$\frac{dp}{dt} = cp(1-p) - ep$$

Levin's (1969) metapopulation model

$$p^* = 1 - \frac{e}{c}$$

Equilibrium metapopulation occupancy

• A. Patch occupancy models

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Levin's (1969) metapopulation model

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Equilibrium metapopulation occupancy

*Note: you have seen this model yesterday. Noticed?

• A. Patch occupancy models

$$\frac{dp}{dt} = c(.)(1-p) - e(.)p$$

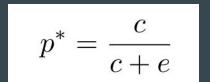
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- The extinction rate e(.) is just a constant: e(.) = e
- The colonization rate is also just a constant: c(.) = c

you get

$$\frac{dp}{dt} = c(1-p) - ep$$

McArthur & Wilson's (1967) continent-island model



Equilibrium island occupancy

• A. Patch occupancy models

Levins' (1969) formed the basis of metapopulation theory MW's (1967) model formed the basis of the theory of island biogeography (TIB)

In both cases:

Colonization should depend on isolation (**distance to mainland** or **patch isolation**): nearer islands/more connected patches are more likely to received migrants

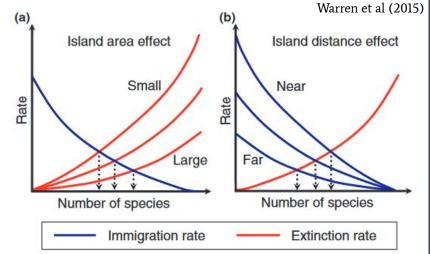
Extinction should depend on island or patch **size** (area): larger islands/patches, having larger population sizes, have lower extinction risk

• A. Patch occupancy models

With several species, we have one (independent) equation per species, and we must sum these equations to get the dynamics of the **number** of species. Take the MW's model.

If all species were similar and non interacting, we'd get **linear** changes in the rates of extinction and colonizations with species richness

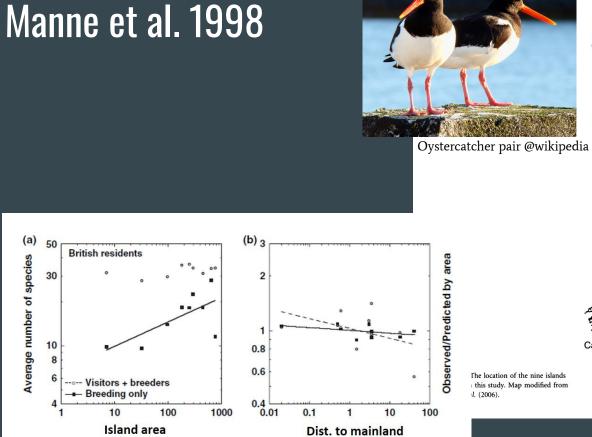
In practice, we rather expect* **convex** functions: – colonization decreases slower than linearly – extinction increases faster than linearly

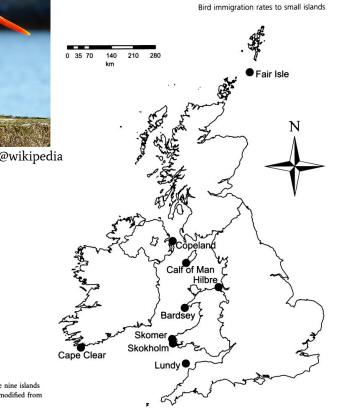


• A. Patch occupancy models

We can **adjust** the MW model to time series of colonizations and extinctions events on islands to test for this prediction: the example of Manne et al. (1998) JAE

The data: presence/absence of breeding birds on 13 British islands, followed over consecutive years (as many as 20 years)



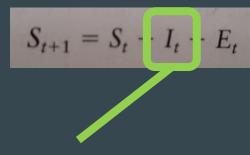


Stracey & Pimm (2009) JoB

The number of species on a particular island at time t (S_t) is known exactly, and modelled as

$$S_{t+1} = S_t + I_t - E_t$$

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Number of immigration events. Probability:

$$\gamma_t = \beta_1 (1 - S_t/P)^{\beta_2}$$

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Number of extinction events. Probability:

$$\theta_t = \beta_3 (S_t/P)^{\beta_4}$$

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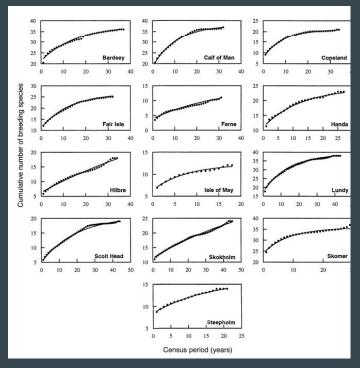
$$\theta_t = \beta_3 (S_t / P)^{1/4}$$

Number of species in regional pool

Manne et al. 1998

From there, we need to:

1. Estimate the number of species in the regional pool(P)



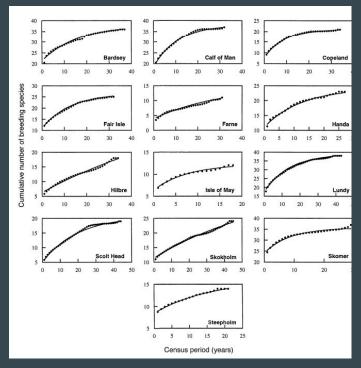
Manne et al. 1998

From there, we need to:

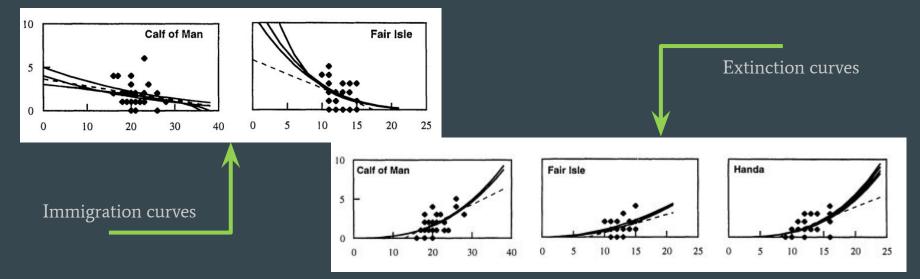
1. Estimate the number of species in the regional pool(P)

2. Express and maximize the likelihood

$$L(\mathbf{I},\mathbf{E};\boldsymbol{\beta}) = \prod_{t=1}^{T} Bin(I_t | P - S_t, \boldsymbol{\gamma}_t) \prod_{t=1}^{T} Bin(E_t | S_t, \theta_t)$$
$$= \prod_{t=1}^{T} \boldsymbol{\gamma}_t^{I_t} (1 - \boldsymbol{\gamma}_t)^{P-S_t-I_t} \theta_t^{E_t} (1 - \theta_t)^{S_t-E_t}$$



For a majority of islands, Manne et al. (1998) did find **non-linear curves**, as expected from theory*.



Manne et al. 1998

*What does it tell us already?

• A. Patch occupancy models

The idea that colonization rate is a function of distance and patch extinction rate varies with patch size can also be incorporated in Levins' metapopulation model (**spatially realistic models**): These are called **incidence function models** (IFM), starting from Hanski (1991)

 $c_i = m_0 \exp\left(-\beta_1 D_i\right)$

 $e_i = m_0 \exp\left(-\beta_2 A_i\right)$

• A. Patch occupancy models

The idea that colonization rate is a function of distance and patch extinction rate varies with patch size can also be incorporated in Levins' metapopulation model (**spatially realistic models**): These are called **incidence function models** (IFM), starting from Hanski (1991)

> Decay rate of colonization with distance (Di)

Decay rate of extinction with patch area (Ai)

• A. Patch occupancy models

IFM models models can be fitted to (snapshot) occupancy data:

- Some parameters have to be estimated separately from other data (e.g., decay rates)
- We must assume dynamic equilibrium for fit, but can be used for projection



American Pika (Moilanen et al. 1998)

Melitaea cinxia (Wahlberg et al. 1996)

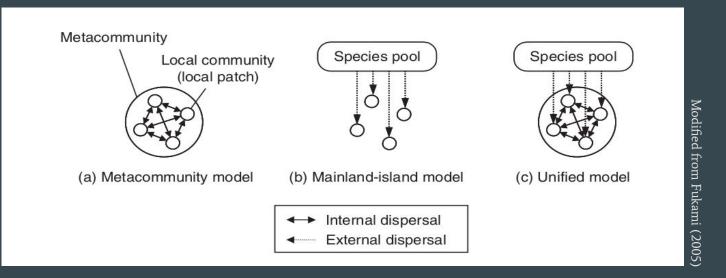


Hyla arborea (Vos et al. 2000)



• A. Patch occupancy models

Of course, you can mix the two assumptions for colonization and obtain a "unified' model*:



• A. Patch occupancy models

Patch occupancy models can be extended in many directions:

- multiple species with interactions (e.g. competition colonization trade-offs)
- different classes of patches (different habitats)
- food webs (trophic chains or networks)

Of course, the more you complicate them, the more difficult analysis is, and also, the more difficult it is to parametrize them and fit them to data.

• A. Patch occupancy models

Even complex patch occupancy models can be fitted to temporal data (time series) - using the temporal turnover (i.e. apparent colonization and extinctions)

This is often called **dynamic occupancy modelling** (MacKenzie, 2003; Bailey et al. 2014)

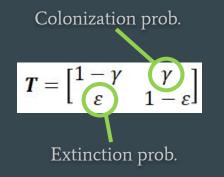
They usually include a detection layer (i.e. a statistical modelling of the observation process)

A. Patch occupancy models

> The idea: modelling patch states (e.g., $\mathbf{X} = \{0, 1\}$) and transition probabilities (matrix \mathbf{T}) among states, such that $\mathbf{X}_{i,t+1} = \mathbf{X}_{i,t}\mathbf{T}$

Often based on Levins' metapopulation model (in discrete time)

equivalent to C(.)



$$P(X_{t+1} = 0) = P(X_t = 0) (1 - \gamma) + P(X_t = 1) \varepsilon$$
$$P(X_{t+1} = 1) = P(X_t = 0) \gamma + P(X_t = 1) (1 - \varepsilon)$$

equivalent to E(.)

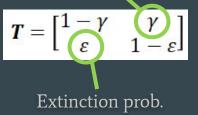
• A. Patch occupancy models

The idea: modelling patch states (e.g., $\mathbf{X} = \{0, 1\}$) and transition probabilities (matrix **T**) among states, such that $\mathbf{X}_{i,t+1} = \mathbf{X}_{i,t}\mathbf{T}$

Often based on Levins' metapopulation model (in discrete time)

equivalent to C(.)

Colonization prob.



$$P(X_{t+1} = 0) = P(X_t = 0) (1 - \gamma) + P(X_t = 1) \varepsilon$$
$$P(X_{t+1} = 1) = P(X_t = 0) \gamma + P(X_t = 1) (1 - \varepsilon)$$

Colonization and extinction can be patch and/or time specific, and then can include patch environmental characteristics, other species effects as well as distance-based dispersal

equivalent to E(.)

One example: metapopulation dynamics of *D. depressissimum* in Guadeloupe (Lamy et al. 2013)

Transition matrix

$$P_{WW} = \begin{bmatrix} 0 & 1\\ 1 - \gamma_W & \gamma_W \\ 1 \left[(1 - \Phi_W) * (1 - \gamma_W) & \Phi_W + (1 - \Phi_W) * \gamma_W \right]$$

With $\Phi = 1 - \varepsilon$; i.e. the persistence probability



One example: metapopulation dynamics of *D. depressissimum* in Guadeloupe (Lamy et al. 2013)

						Estin	Estimates in wet sites (CI)		
			Model	q _{i,t}	Parameter	Pw	Φ_{W}	$\gamma_{\rm W}$	
Table 4: Posterior mea their 95% credible into		ot and covariate effects in the		+	$\Phi_{\rm W} + \gamma_{\rm W}$.79 (.75, .82)	.76 (.69, .82)	.32 (.28, .36)	
Parameter	Coefficient	Mean (95% CI)	Predicted mean		Detection	ah			
$\Phi_{W_{Lt}}$.787		Detection pr	od. Mean pe	ersistence pro	b	
Intercept	a_1	1.31 (.883, 1.758)				*	*		
Vegetation	β_2	.961 (.650, 1.297)					Mean coloniz	zation prob.	
Little rainy season	β_4	631 (998,298)							
$\gamma_{W_{tr}}$.288						
Intercept	a_3	843 (-1.100,599)							
Size	B ₉	.382 (.166, .604)			Envt. effects	s on persistenc	e		
Stability	β_{11}	839 (-1.180,538)	R						
Connectivity	β_{12}	.353 (.142, .575)							
pw:	12010555								
Detectability		.777 (.742, .811)			Envt effect	ts on colonizat	ion		

E d'and in the (CI)

Note: Each coefficient is given in logit scale. Consequently, for each parameter (Φ_W , Φ_D , γ_W , and γ_D), we computed the predicted mean by setting all covariates to their conditional means in wet sites (for Φ_W and γ_W) or dry sites (for Φ_D and γ_D) in the prediction formula (eqq. [4]–[7]) and using the inverse-logit function to back-transform into the natural scale.

With multiple species: *P. acuta & A. marmorata -* Guadeloupe

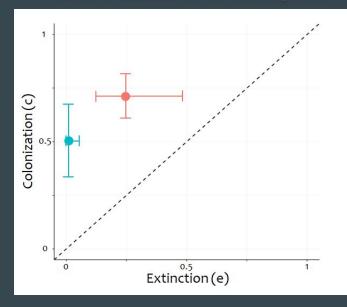
Rates in **absence** of the other species



Aplexa marmorata

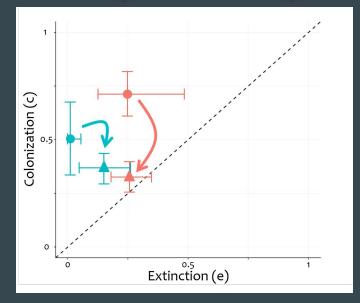


Physa acuta



With multiple species: *P. acuta & A. marmorata -* Guadeloupe

Rates in **presence** of the other species





Aplexa marmorata



Physa acuta

A. marmorata increases P. acuta extinction rate

P. acuta reduces A. marmorata colonization rate

With multiple species: *P. acuta & A. marmorata -* Guadeloupe

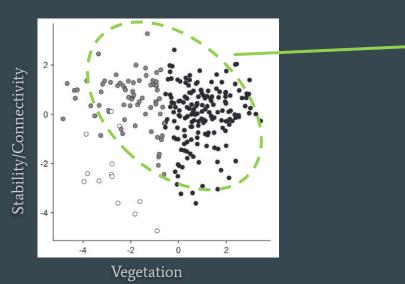
These interaction effects depends on environmental conditions



Aplexa marmorata



Physa acuta



A. marmorata fundamental niche

With multiple species: *P. acuta & A. marmorata -* Guadeloupe

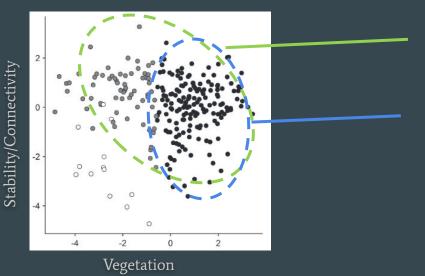
These interaction effects depends on environmental conditions



Aplexa marmorata



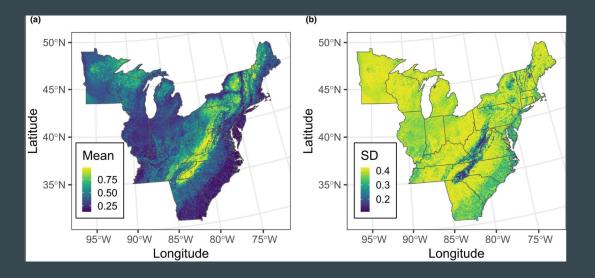
Physa acuta



A. marmorata fundamental niche

A. marmorata realized niche

There are increasingly flexible and easy-to-use solutions to implement occupancy models



• B. Introducing local population dynamics

All previous mechanistic models describe presence/absence only (i.e. neglected local dynamics)

It may be more realistic to describe within patch population dynamics, and more data could be used (when abundance data is available)

• B. Integrating local population dynamics

> The simplest approach possible is to take a patch occupancy model (e.g. Levins' model), and add a dynamical variable to describe local population density (e.g. logistic growth within each patch). This approach was taken for instance by Gyllenberg & Hanski (1992)

> Alternatively, one can simply distinguish "small" populations (i.e. recently colonized) from "large" populations (small populations that have grown). This adds a second equation to Levins' model, but remains quite simple (Hanski 1985)

This can cause **rescue effects** and **alternative stable states** (similar to an Allee effect at metapopulation level). It has implications for the response to perturbations (Eriksson et al. 2014)

• B. Integrating local population dynamics

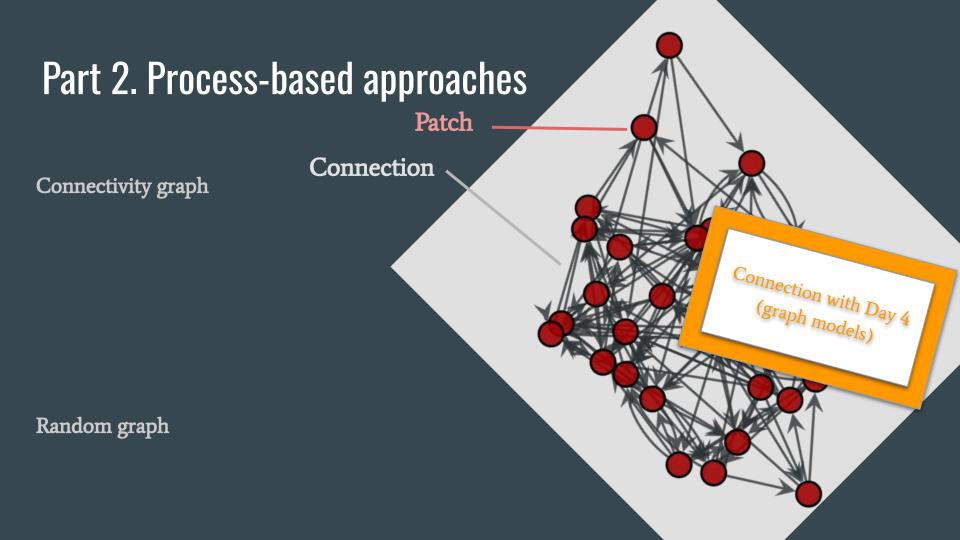
There are, again, many models that attempt to do this. > Another example is Hubbell's **neutral model of biodiversity** over several connected communities.

Owing to the fact that all species are assumed identical, the dynamics of local abundances (assuming some constant total abundance) is relatively easy to model.

Etienne (2<u>005</u>)

$$P[D|\theta, m, J] = \frac{J!}{\prod_{i=1}^{S} n_i \prod_{j=1}^{J} \Phi_j!} \frac{\theta^S}{(\theta)_J}$$
$$\times \sum_{\mathcal{A}=S}^{J} \left(K(D, \mathcal{A}) \frac{(\theta)_J}{(\theta)_{\mathcal{A}}} \frac{I^{\mathcal{A}}}{(I)_J} \right)$$

Part 2. Process-based approaches Patch Connectivity graph Connection Fully connected (complete)



Connectivity graph

Planar graph

• C. Models with localized dispersal

Lattice (grid) graph

• C. Models with localized dispersal

> **Reaction-diffusion models** are a convenient framework to model the spatial spread of species, combining local dispersal (diffusion) and local population dynamics (reaction)

$$\begin{cases} \partial_t U(t,x) &= \Delta D(x,\Theta)U(t,x) + F(\Theta,U(t,x)) \\ U(0,x) &= U_0(x,\Theta) \\ \vec{\nabla} D(x,\Theta)U(t,x).\,\vec{n} &= 0 \end{cases}$$

• C. Models with localized dispersal

> **Reaction-diffusion models** are a convenient framework to model the spatial spread of species, combining local dispersal (diffusion) and local population dynamics (reaction)

$$rac{\partial u}{\partial t} - D rac{\partial^2 u}{\partial x^2} = ru(1-u).$$

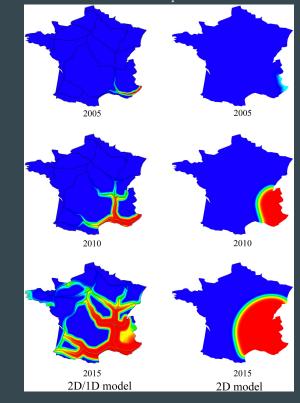
Fisher KPP in 1D

• B. Models with local population dynamics

Can be simulated and fitted to data (Roques & Bonnefond 2016)



Roques & Bonnefon 2016



• B. Models with local population dynamics

Can be extended to several interacting species

$$\begin{cases} \partial_t u - \partial_{xx} u = u \left(1 - u - av \right) & \text{in } (0, +\infty) \times \mathbb{R} \\ \partial_t v - d\partial_{xx} v = rv \left(1 - v - bu \right) & \text{in } (0, +\infty) \times \mathbb{R} \\ u \left(0, x \right) = \tilde{u}_0 \left(x \right) & \text{for all } x \in \mathbb{R} \\ v \left(0, x \right) = 1 - \tilde{v}_0 \left(x \right) & \text{for all } x \in \mathbb{R} \end{cases}$$

Girardin & Lam (2010) (two species-competition-diffusion)

Predator-prey diffusion

• B. Models with local population dynamics

Describing the local population dynamics with several interacting species quickly becomes **too complicated** to be usefully modelled with simple mathematical models.

In this case, we usually turn to **stochastic simulation models** .

• C. Stochastic simulation models

Various environments are now available to simulate spatial data



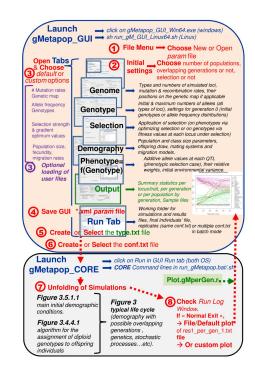
Ecological Modelling Volume 492, June 2024, 110730



MetaIBM: A Python-based library for individual-based modelling of ecoevolutionary dynamics in spatial-explicit metacommunities

Jian-Hao Lin ª, Yu-Juan Quan ^b, Bo-Ping Han ª 옷 図

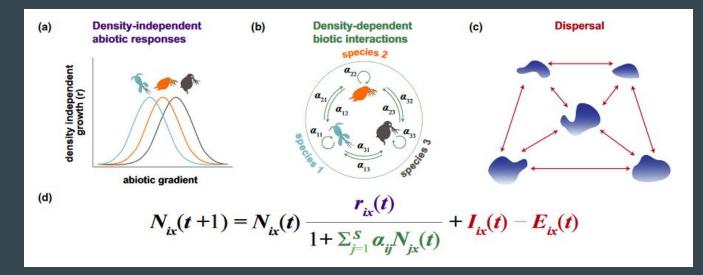
MetaIBM (2024)



gMetapop (2024)

• C. Stochastic simulation models

One relatively general metacommunity model has been proposed by Thompson et al. (2020)



Stochasticity is modelled as the random draw of the values of E(t) and N(t)

Conclusion

Ordination methods

Pros:

- Very large litterature
- Widely used/applicable
- Able to work on (very) large datasets

Cons:

- Links with theory unclear
- So many methods
- Sometimes controversial

JSDMs

Pros:

- Sophisticated models with will for linking components to processes
- Very flexible
- Can assess relationship between species traits/phylo. and 'niche'

Cons:

- Still SDMs with residual correlation as criterion
- Results interpretation not always clear

Process-based models

Pros:

- Explicit processes and link to theory
- Can be customized to specific system based on knowledge
- Can address 'what if' questions

Cons:

- Greedy in data & comp. ressources
- Limited in the number of species they can handle
- Highly parametric

Conclusion

• So what?

Conclusion

• So what?

So, no good method!

"In conclusion, we find no reason why a proponent of either of the two extremes of correlative and process-based species distribution modelling should hold the moral high ground.' Correlationists' should be humble: their model's success maybe due to spurious correlations. 'Mechanists' should be unassertive about their approach, because they will only find effects of processes that they included. Either approach must comply with nature, statistically or mechanistically, and be aware of the kinds of questions they are best suited to answer." Dormann et al. 2012

To go further

Main reference textbooks:

Legendre & Legendre. **2012**. *Numerical Ecology*.

Vellend. 2016. The Theory of Ecological Communities.

Leibold & Chase. 2018. Metacommunity Ecology.

Mittelbach & McGill. 2019. Community Ecology.

Ovaskainen & Abrego. 2020. Joint Species Distribution Modelling

Ready for practicals?

• questions on morning session



Afternoon: practicals

1. Simulate data/test methods

Simulate with the following code:

https://github.com/mxdub/TMDSpatial

Try different analyses \rightarrow Does it work?

To install : JAGS R packages : vegan, R2jags, Hmsc And devtools::install_github('mxdub/TMDSpatial')

2. The effects of patch disturbance

Article walkthrough:

Functional Ecology



RESEARCH ARTICLE 👌 Open Access 💿 🛈

Metacommunity dynamics and the detection of species associations in co-occurrence analyses: Why patch disturbance matters

Vincent Calcagno, Nik J. Cunniffe 🔀 Frédéric M. Hamelin

First published: 04 April 2022 | https://doi.org/10.1111/1365-2435.14047

Go fetch the R markdown at: <u>https://github.com/nikcunniffe/MetacommunityDynamics</u>

Practical 2: the effects of patch disturbance

- 1. What is patch disturbance?
- 2. How do you include it in Levins' metapopulation model?(write down the equations)(walk through model formulation: pach age, equilibrium...)
- 3. What consequences for species co-occurrence patterns?
- 4. Test the expectations/predictions